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Effect of Resource Dynamics on Species Packing in Diverse Ecosystems

Wenping Cui^{®*}

Department of Physics, Boston University, 590 Commonwealth Avenue, Boston, Massachusetts 02139, USA and Department of Physics, Boston College, 140 Commonwealth Avenue, Chestnut Hill, Massachusetts 02467, USA

Robert Marsland III^{®†} and Pankaj Mehta^{®‡}

Department of Physics, Boston University, 590 Commonwealth Avenue, Boston, Massachusetts 02139, USA

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The competitive exclusion principle asserts that coexisting species must occupy distinct ecological niches (i.e., the number of surviving species cannot exceed the number of resources). An open question is to understand if and how different resource dynamics affect this bound. Here, we analyze a generalized consumer resource model with externally supplied resources and show that—in contrast to self-renewing resources—species can occupy only half of all available environmental niches. This motivates us to construct a new schema for classifying ecosystems based on species packing properties.

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One of the most stunning aspects of the natural world is the incredible diversity of species present in many environments [1,2]. A major goal of community ecology is to understand the rules governing community structure and species coexistence patterns in these complex ecosystems. One promising approach that has recently emerged for tackling this challenge is to use ideas from statistical mechanics inspired by spin glass physics [3,4]. In such an approach, ecosystems are viewed as large interacting disordered systems, allowing for the identification of universal, collective properties [5,6]. Such statistical physics inspired models are also able to reproduce many experimental observations, especially in the context of microbial ecosystems [7–9].

Much of this work has focused on generalized Lotka-Volterra models where species directly interact with each other in a pairwise fashion [5,10-16]. While such models have led to deep ecological insights [17] and have allowed for the identification of interesting ecological phases and phase transitions [10-12], a major drawback of Lotka-Volterra models are that they do not explicitly model the resources present in the ecosystem. Instead, resource dynamics are implicitly represented through the choice of species-species interactions making it difficult to understand the relationship between resource dynamics and community structure.

In contrast, generalized consumer resource models (GCRMs), first introduced by MacArthur and Levins in a series of seminal papers [18–20], explicitly incorporate both species and resource dynamics. In GCRMs, ecosystems are described by species that can consume and deplete resources according to a set of consumer preferences. Interactions between species arise because species with similar consumer preferences occupy similar environmental niches and hence compete for common resources. An important theoretical

and conceptual result that follows from GCRMs is that the number of species that can coexist in an ecosystem is limited by the number of resources that are present. In other words, if we denote the number of species that can survive in an ecosystem by S^* and the number of supplied resources as M, the competitive exclusion principle yields an upper bound for the amount of species that can be packed into the ecosystem: $(S^*/M) \leq 1$ [21].

The basic intuition behind this bound is that the growth rates $q_i(\mathbf{R})$ of all coexisting species $i = 1, 2, \dots$ must simultaneously vanish, and since the space of resource concentrations \mathbf{R} is M dimensional, at most M of these equations can be simultaneously solved [see Supplemental Material (SM) for discussion of nongeneric phenomena where the bound is violated [22]]. While this result gives an upper bound, it is not clear when and if it will be saturated. In particular, we show below that the choice of resource dynamics fundamentally alters species packing properties. To show this, we analyze GCRMs with two different resource dynamics: self-renewing resources where resources grow logistically in the absence of consumers [18,19] and externally supplied resources that are supplied and degraded at a constant rate [23–25] (see Fig. 1). We derive species packing bounds for both choices of dynamics by analyzing the susceptibilities of a new cavity solution for GCRMs with externally supplied resources and combining it with the previously derived cavity solution for GCRMs with self-renewing resources [6,26,27]. Surprisingly, in the absence of metabolic trade-offs we find that, for externally supplied resources, species can occupy only half of all available resource niches: $(S^*/M) < \frac{1}{2}$. Motivated by these results, we suggest a new schema for classifying ecosystems based on their species packing properties.

Model.—GCRMs describe the ecological dynamics of *S* species of consumers N_i (i = 1, 2, ..., S) that can consume *M* distinct resources R_{α} ($\alpha = 1, 2, ..., M$). The rate at which species N_i consumes and depletes resource R_{β} is encoded in a matrix of consumer preferences $C_{i\beta}$. In order to survive, species have a minimum maintenance cost m_i . Equivalently, m_i can also be thought of as the death rate of species *i* in the absence of resources. These dynamics can be described using a coupled set of M + S ordinary differential equations of the form

$$\frac{dN_i}{dt} = N_i \sum_{\beta} C_{i\beta} R_{\beta} - N_i m_i,$$

$$\frac{dR_{\alpha}}{dt} = h_{\alpha}(R_{\alpha}) - \sum_j N_j C_{j\alpha} R_{\alpha},$$
 (1)

where $h_{\alpha}(R_{\alpha})$ is a function that describes the dynamics of the resources in the absence of any consumers (see Fig. 1).

For self-renewing resources (e.g., plants, animals), the dynamics can be described using logistic growth of the form

$$h_{\alpha}(R_{\alpha}) = R_{\alpha}(\kappa_{\alpha} - R_{\alpha}), \qquad (2)$$

with κ the carrying capacity. While such resource dynamics is reasonable for biotic resources, abiotic resources such as minerals and small molecules cannot self-replicate and are usually supplied externally to the ecosystem [Fig. 1(b)]. A common way to model this scenario is by using linearized resource dynamics of the form

$$h_{\alpha}(R_{\alpha}) = K_{\alpha} - \omega_{\alpha} R_{\alpha}.$$
 (3)



FIG. 1. Schematic description for two types of resources. (a) Self-renewing resources (e.g., plants), which are replenished through organic reproduction, and (b) externally supplied resources (e.g., nutrients that sustain gut microbiota), which are replenished by a constant flux from some external source, and diluted at a constant rate. (c) The supply rate as a function of resource abundance for both choices, with $\kappa = \omega_{\alpha} = K_{\alpha} = 1$.

Figure 1(c) shows a plot of these two choices. Note that the two resource dynamics behave very differently at low resource levels. The self-renewing resources can go extinct and eventually disappear from the ecosystem, while this is not true of externally supplied resources.

Recent research has shown some unexpected and interesting nongeneric phenomena can appear in GCRMs in the presence of additional constraints on parameter values. A common choice of such constraints is the imposition of a "metabolic budget" on the consumer preference matrix [23,28] tying the maintenance cost m_i to the total consumption capacity $\sum_{\beta} C_{i\beta}$ [24,29]. These metabolic tradeoffs can be readily incorporated into the cavity calculations and have significant impacts on species packing, as will be discussed below.

Cavity solution.-Recently, we derived a mean-field cavity solution for steady-state dynamics of the GCRM with self-renewing resource dynamics in the high-dimensional limit where the number of resources and species in the regional species pool is large $(S, M \gg 1)$ [6,26,27]. The overall procedure for deriving the cavity equations for GCRM with externally supplied resource is similar to that for GCRMs with self-renewing resources and is shown in Fig. S1 in the SM [22]. We assume the K_{α} and m_i are independent random normal variables with means K and mand variances σ_K^2 and σ_m^2 , respectively. We also assume ω_{α} are independent normal variables with mean ω and variance σ_{ω}^2 . The elements of the consumption matrix $C_{i\alpha}$ are drawn independently from a normal distribution with mean μ/M and variance σ_c^2/M . This scaling with M is necessary to guarantee that $\langle N \rangle$, $\langle R \rangle$ do not vanish when S, $M \gg 1$ with $M/S = \gamma$ fixed. Later, we will consider a slightly modified scenario where the maintenance costs are correlated with the consumption matrix in order to implement the metabolic trade-offs discussed above.

The basic idea behind the cavity method is to derive self-consistency equations relating an ecosystem with M resources and S species to an ecosystem with M + 1 resources and S + 1 resources. This is done by adding a new "cavity" species 0 and a new "cavity" resource 0 to the original ecosystem. When $S, M \gg 1$, the effect of the new cavity species or resource is small and can be treated using perturbation theory. The cavity solution further exploits the fact that since the $C_{i\alpha}$ are random variables, when $M \gg 1$ the sum $\sum_{\alpha} C_{i\alpha} R_{\alpha}$ will be well described by a normal distribution with mean $\mu \langle R \rangle$ and variance $\sigma_c^2 q_R$, where $q_R = \langle R^2 \rangle = 1/M \sum_{\alpha} R_{\alpha}^2$ (see SM for details [22]). Combining this with the non-negativity constraint, the species distribution can be expressed as a truncated normal distribution,

$$\bar{N} = \max\left[0, \frac{\mu \langle R \rangle - m + \sqrt{\sigma_c^2 q_R + \sigma_m^2} z_N}{\sigma_c^2 \chi}\right], \quad (4)$$

where $\chi = -\langle \partial \bar{R}_{\alpha} / \partial \omega_{\alpha} \rangle = -M^{-1} \sum_{\alpha} (\partial \bar{R}_{\alpha} / \partial \omega_{\alpha})$ and z_N is a standard normal variable. This equation describes

GCRMs with both externally supplied and self-renewing resource dynamics [26].

The steady-state cavity equations for externally supplied resources are significantly more complicated and technically difficult to work with than the corresponding equations for self-renewing resources. To see this, note that the steady-state abundance of resource α can be found by plugging Eq. (3) into Eq. (1) and setting the left-hand side to zero to get

$$\bar{R}_{\alpha} = K_{\alpha} / \left(\omega_{\alpha} + \sum_{j} \bar{N}_{j} C_{j\alpha} \right) = \frac{K_{\alpha}}{\omega_{\alpha}^{\text{eff}}}, \quad (5)$$

where we have defined $\omega_{\alpha}^{\text{eff}} = \omega_{\alpha} + \sum_{j} \bar{N}_{j} C_{j\alpha}$. When $S \gg 1$, both the denominator $\omega_{\alpha}^{\text{eff}}$ and the numerator K_{α} can be modeled by independent normal random variables. This implies that the steady-state resource abundance is described by a ratio of normal variables (i.e., the normal ratio distribution) instead of a truncated Gaussian as in the self-renewing case [30] (see Fig. S5 of SM [22]). At large σ_c , this makes solving the cavity equations analytically intractable. Luckily, if the variance of the denominator $\omega_{\alpha}^{\text{eff}}$ is small compared with the mean—which is true when σ_c not too large-we can still obtain an approximate replicasymmetric solution by expanding in powers of the standard deviation over the mean of $\omega_{\alpha}^{\text{eff}}$ (see SM [22]). We consider expansions to the cavity solutions where the denominator in Eq. (5) is expanded to first order. In general, the backreaction correction is quite involved since resources and species form loopy interactions resulting in nontrivial correlation between $C_{i\alpha}$ and N_i that must be properly accounted for (see SM [22]).

Comparison with numerics.—The full derivation of firstorder expansions of the mean-field equations is given in the SM [22]. The resulting self-consistency equations can be solved numerically in *Mathematica*. Figure 2 shows a comparison between the cavity solution and 1000 independent numerical simulations for various ecosystem properties such as the fraction of surviving species S^*/S and the first and second moment of the species and resource distributions (simulation details are in the SM [22]). As can be seen in the figure, our analytic expressions agree remarkably well over a large range of σ_c . However, at very large σ_c (not shown), the cavity solutions start deviating from the numerical simulations because the ratio normal distribution can no longer be described using the first- order expansion to the full cavity equations.

As a further check on our analytic solution, we ran simulations where the $C_{i\alpha}$ were drawn from different distributions. One pathology of choosing $C_{i\alpha}$ from a Gaussian distribution is that when σ_c is large, many of consumption coefficients are negative. To test whether our cavity solution still describes ecosystems when $C_{i\alpha}$ are strictly positive, we compare our cavity solution to simulations where the $C_{i\alpha}$ are drawn from a Bernoulli or uniform



FIG. 2. Comparison between cavity solutions (see text for definition) and simulations for the fraction of surviving species $\phi_N = (S^*/S)$ and the first and second moments of the species and resources distributions as a function of σ_c . The error bar shows the standard deviation from 1000 numerical simulations with M = S = 100, and all other parameters are defined in the SM [22]. Simulations were run using the CVXPY package [31].

distribution. As before, there is remarkable agreement between analytics and numerics (see Fig. S2 [22]).

Species packing without metabolic trade-offs.—The essential ingredients needed to derive species packing bounds for GCRMS are the cavity equations for the average local susceptibilities, $\nu = \langle \partial \bar{N}_i / \partial m_i \rangle = S^{-1} \sum_j (\partial \bar{N}_i / \partial m_i)$ and $\chi = \langle \partial \bar{R}_\alpha / \partial X_\alpha \rangle = M^{-1} (\partial \bar{R}_\alpha / \partial X_\alpha)$, with $X_\alpha = K_\alpha$ for externally supplied resources and $X_\alpha = -\omega_\alpha$ for selfrenewing resources. These two susceptibilities measure how the mean species abundance and mean resource abundance respond to changes in the species death rate and the resource supply (depletion) rate, respectively. They play an essential role in the cavity equation and can be used for distinguishing different phases in complex systems [6,32].

For the self-renewing case, the susceptibilities χ_s and ν_s are given by Eqs. (59) and (60) in Ref. [27],

$$\nu_s = -\frac{\phi_N}{\sigma_c^2 \chi_s}, \qquad \chi_s = \frac{\phi_R}{1 - \gamma^{-1} \sigma_c^2 \nu_s}, \tag{6}$$

and can be reduced to $\chi_s = \phi_R - \gamma^{-1} \phi_N$, where $\phi_R = M^*/M$, with M^* equal to the number of nonextinct resources in the ecosystem. In order to guarantee the positivity of $\langle N \rangle$, we must have $\chi_s = \phi_R - \gamma^{-1} \phi_N > 0$, resulting in an upper bound,

$$1 \ge \frac{M^*}{M} > \frac{S^*}{M},\tag{7}$$

which states that the number of surviving resources must be larger than the number of surviving species.



FIG. 3. Comparison of the species packing ratio S^*/M at various σ_c and K for self-renewing and externally supplied resource dynamics. The simulations represent averages from 1000 independent realizations with the system size M = 100, S = 500 (parameters in SM [22]).

For the externally supplied case, the corresponding equations take the form

$$\nu = -\frac{\phi_N}{\sigma_c^2 \chi}, \qquad \chi = -\frac{1}{2\gamma^{-1}\nu\sigma_c^2}(1 - \langle \cdots \rangle), \qquad (8)$$

where the full expression of $\langle \cdots \rangle$ can be found in Eq. (63) in the SM [22]. For our purposes, the most important property is that in the absence of metabolic trade-offs, the expression $\langle \cdots \rangle$ is always *positive*. Combining this observation with the equations above gives the upper bound:

$$\frac{1}{2} > \frac{S^*}{M} = \phi_N \gamma^{-1}.$$
 (9)

Thus, for externally supplied resources, at most half of all potential niches are occupied. Figure 3 shows numerical simulations confirming the species packing bound for various choices of *K* and σ_c (see Fig. S6 in SM for various choices of *S/M* [22]). The lower diversity found when resources are supplied externally can be anticipated by noting that the resource abundance in this model is more narrowly distributed than in a model with self-renewing resources. As a result, species experience stronger competition (see Fig. S5 and more details in SM [22]). However, we still currently lack an intuitive explanation of why the species packing bound is exactly 0.5.

Species packing with metabolic trade-offs.—We also find that metabolic trade-offs modify the cavity equations in such a way that the expression in brackets $\langle \cdots \rangle$ in Eq. (8) can become negative (see SM [22]). However, it still remains greater than -1, allowing us to derive a species packing bound of the form $S^* < M$ even in the presence of soft metabolic constraints. In Fig. 4, we simulated various ecosystems where the maintenance costs of species were chosen to obey metabolic trade-offs of the form $m_i = \sum_{\alpha} C_{i\alpha} + \delta m_i$, where δm_i are independent and identically distributed (i.i.d.) normal variables with variance σ_m^2 . Note that a larger σ_m corresponds to ecosystems with softer metabolic constraints. We found that when $\sigma_m/\sigma_c > 1$, these



FIG. 4. Species packing bounds in the presence of metabolic trade-offs. (a) The species packing ratio S^*/M as a function of σ_m/σ_c , where σ_m is the standard deviation of the δm_i and σ_c/\sqrt{M} is the standard deviation of $C_{i\alpha}$. Simulations are for binary consumer preference matrix $C_{i\alpha}$ drawn from a Bernoulli distribution with probability *p*. (b) m_i versus $\sum_{\alpha} C_{i\alpha}$ for p = 0.1 and $\sigma_m/\sigma_c = 10^{-0.5}$. See SM for all parameters [22].

ecosystems obey the 1/2 species packing bound derived above. This can also be analytically shown using the modified cavity equations derived in the SM [22]. Finally, we show in the SM that when the metabolic trade-offs take the form of hard constraints on the consumer preferences as in Refs. [23,24,28,29], the cavity equations allow for interesting nongeneric behavior with $S^* \ge M$, consistent with these previous works. Importantly, we find that even modest modifications of the trade-off equation $m_i \propto \sum_{\alpha} C_{i\alpha}$ result in ecosystems that satisfy the 1/2 species packing bound.

Classifying ecosystems using species packing.— Recently, it has become clear that there is a deep relationship between ecosystem and constraint satisfaction problems [24,25,27,29]. In particular, each species can be thought of as a constraint on possible resource abundances [25,27]. Inspired by jamming [33], this suggests that we can separate ecosystems into qualitatively distinct classes depending on whether the competitive exclusion bound is saturated. We designate ecosystems where $S^* \rightarrow M$ (like GCRMs with self-renewing resources) as isostatic species packings, and ecosystems where the upper bound S_{max} on the number of surviving species is strictly less than the number of resources $S^* < S_{\text{max}} < M$ (like GCRMs with externally supplied resources without metabolic trade-offs) as hypostatic species packings. Ecosystems with $S^* \ge M$ (like GCRMs with hard metabolic constraints) are designated as nongeneric species packings because of the presence of a macroscopic number of additional hard constraints (i.e., the number of additional constraints that are imposed scales with S and M in the limit $S, M \to \infty$). This basic schema suggests a way of refining the competitive exclusion principle and may help shed light on controversies surrounding the validity of basic species packing bounds.

Discussion.—In this Letter, we examine the effect of resource dynamics on community structure and large-scale ecosystem level properties. To do so, we analyzed

generalized consumer resource models with two different resource dynamics: externally supplied resources that are supplied and degraded at a constant rate and self-replicating resources whose behavior in the absence of consumers is well described by a logistic growth law. Using a new cavity solution for GCRMs with externally supplied resources and a previously found cavity solution of the GCRM with self-renewing resources, we show that the community structure is fundamentally altered by the choice of resource dynamics. In particular, for externally supplied resources, we find that species generically can only occupy *half* of all available niches, whereas for self-renewing resources all environmental niches can be filled. We confirm this surprising bound using numerical simulations.

In this Letter, we consider the effect of metabolic tradeoffs and show that they can increase species packing in an ecosystem. In the future, it will be interesting to ask how other specialized network structures, including niche partitioning, higher specialization, or combinations of specialists and generalists, can affect our results. Based on our experience, we expect that, even in these more complicated ecosystems, our species packing bound will hold quite generically. But much more work needs to be done to confirm if this is really the case.

Our results show how resource dynamics, which are neglected in commonly used Lotka-Volterra models, can fundamentally alter the properties of ecosystems. Much work still needs to be done to see if and how our results must be modified to account for other ecological processes such as demographic stochasticity, spatial structure, and microbe-specific interactions such as cross feeding [7,8]. It will also be necessary to move beyond steady states and consider the dynamical properties of these ecosystems. More generally, it will be interesting to further explore the idea that we can classify ecosystems based on species packing properties and see if such a schema can help us better understand the origins of the incredible diversity we observe in real-world ecosystems.

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^{*}cuiw@bu.edu [†]marsland@bu.edu [‡]pankajm@bu.edu

- C. Huttenhower, D. Gevers, R. Knight, S. Abubucker, J. H. Badger, A. T. Chinwalla, H. H. Creasy, A. M. Earl, M. G. FitzGerald, R. S. Fulton *et al.*, Structure, function and diversity of the healthy human microbiome, Nature (London) **486**, 207 (2012).
- [2] A. H. Gentry, Tree species richness of upper Amazonian forests, Proc. Natl. Acad. Sci. U.S.A. 85, 156 (1988).

- [3] M. Mézard, G. Parisi, and M. Virasoro, Spin Glass Theory and Beyond: An Introduction to the Replica Method and its Applications (World Scientific Publishing Company, Singapore, 1987), Vol. 9.
- [4] H. Nishimori, Statistical Physics of Spin Glasses and Information Processing: An Introduction (Clarendon Press, Oxford, 2001), Vol. 111.
- [5] M. Barbier, J.-F. Arnoldi, G. Bunin, and M. Loreau, Generic assembly patterns in complex ecological communities, Proc. Natl. Acad. Sci. U.S.A. 115, 2156 (2018).
- [6] W. Cui, R. Marsland III, and P. Mehta, Diverse communities behave like typical random ecosystems, arXiv:1904.02610.
- [7] J. E. Goldford, N. Lu, D. Bajić, S. Estrela, M. Tikhonov, A. Sanchez-Gorostiaga, D. Segrè, P. Mehta, and A. Sanchez, Emergent simplicity in microbial community assembly, Science 361, 469 (2018).
- [8] R. Marsland III, W. Cui, J. Goldford, A. Sanchez, K. Korolev, and P. Mehta, Available energy fluxes drive a transition in the diversity, stability, and functional structure of microbial communities, PLoS Comput. Biol. 15, e1006793 (2019).
- [9] R. Marsland III, W. Cui, and P. Mehta, A minimal model for microbial biodiversity can reproduce experimentally observed ecological patterns, Sci. Rep. 10, 3308 (2020).
- [10] C. K. Fisher and P. Mehta, The transition between the niche and neutral regimes in ecology, Proc. Natl. Acad. Sci. U.S.A. 111, 13111 (2014).
- [11] D. A. Kessler and N. M. Shnerb, Generalized model of island biodiversity, Phys. Rev. E 91, 042705 (2015).
- [12] G. Bunin, Ecological communities with Lotka-Volterra dynamics, Phys. Rev. E 95, 042414 (2017).
- [13] M. Barbier and J.-F. Arnoldi, The cavity method for community ecology, bioRxiv, https://doi.org/10.1101/147728 (2017).
- [14] G. Biroli, G. Bunin, and C. Cammarota, Marginally stable equilibria in critical ecosystems, New J. Phys. 20, 083051 (2018).
- [15] F. Roy, G. Biroli, G. Bunin, and C. Cammarota, Numerical implementation of dynamical mean field theory for disordered systems: application to the Lotka–Volterra model of ecosystems, J. Phys. A 52 484001 (2019).
- [16] M. T. Pearce, A. Agarwala, and D. S. Fisher, Stabilization of extensive fine-scale diversity by ecologically driven spatiotemporal chaos, Proc. Natl. Acad. Sci. U.S.A. 117, 14572 (2019).
- [17] P. Chesson, Mechanisms of maintenance of species diversity, Annu. Rev. Ecol. Syst. 31, 343 (2000).
- [18] R. MacArthur and R. Levins, The limiting similarity, convergence, and divergence of coexisting species, Am. Nat. 101, 377 (1967).
- [19] P. Chesson, MacArthur's consumer-resource model, Theor. Popul. Biol. 37, 26 (1990).
- [20] R. MacArthur, Species packing and competitive equilibrium for many species, Theor. Popul. Biol. **1**, 1 (1970).
- [21] R. McGehee and R. A. Armstrong, Some mathematical problems concerning the ecological principle of competitive exclusion, J. Differ. Equations 23, 30 (1977).
- [22] See Supplemental Material at http://link.aps.org/supplemental/ 10.1103/PhysRevLett.125.048101 for details about numerical simulations and analytical computations.

- [23] A. Posfai, T. Taillefumier, and N. S. Wingreen, Metabolic Trade-Offs Promote Diversity in a Model Ecosystem, Phys. Rev. Lett. 118, 028103 (2017).
- [24] M. Tikhonov and R. Monasson, Collective Phase in Resource Competition in a Highly Diverse Ecosystem, Phys. Rev. Lett. **118**, 048103 (2017).
- [25] R. Marsland III, W. Cui, and P. Mehta, The minimum environmental perturbation principle: A new perspective on Niche theory, arXiv:1901.09673.
- [26] M. Advani, G. Bunin, and P. Mehta, Statistical physics of community ecology: a cavity solution to MacArthur's consumer resource model, J. Stat. Mech. (2018) 033406.
- [27] P. Mehta, W. Cui, C.-H. Wang, and R. Marsland III, Constrained optimization as ecological dynamics with applications to random quadratic programming in high dimensions, Phys. Rev. E 99, 052111 (2019).

- [28] Z. Li, B. Liu, S. H.-J. Li, C. G. King, Z. Gitai, and N. S. Wingreen, Modeling microbial metabolic trade-offs in a chemostat, bioRxiv, 664698 (2019).
- [29] A. Altieri and S. Franz, Constraint satisfaction mechanisms for marginal stability and criticality in large ecosystems, Phys. Rev. E 99, 010401(R) (2019).
- [30] G. Marsaglia *et al.*, Ratios of normal variables, J. Stat. Softw. **16**, 1 (2006).
- [31] A. Agrawal, R. Verschueren, S. Diamond, and S. Boyd, A rewriting system for convex optimization problems, J. Control Decis. 5, 42 (2018).
- [32] M. Ramezanali, P. P. Mitra, and A. M. Sengupta, The cavity method for analysis of large-scale penalized regression, arXiv:1501.03194.
- [33] A. J. Liu and S. R. Nagel, The jamming transition and the marginally jammed solid, Annu. Rev. Condens. Matter Phys. 1, 347 (2010).